# Hybridization among dominant tree species correlates positively with understory plant diversity<sup>1</sup>

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- Premise of the study: Elucidating the factors that determine the abundance and distribution of species remains a central goal of
  ecology. It is well recognized that genetic differences among individual species can affect the distribution and species interactions of dependent taxa, but the ecological effects of genetic differences on taxa of the same trophic level remain much less
  understood. Our goal was to test the hypothesis that differences between related overstory tree species and their hybrids can
  influence the understory plant community in wild settings.
- Methods: We conducted vegetation surveys in a riparian community with the overstory dominated by Populus fremontii, P. angustifolia, and their natural hybrids (referred to as cross types) along the Weber River in north central Utah, USA. Understory diversity and community composition, as well as edaphic properties, were compared under individual trees.
- *Key results*: Diversity metrics differ under the three different tree cross types such that a greater species richness, diversity, and cover of understory plants exist under the hybrids compared with either of the parental taxa (30–54%, 40–48%, and 35–74% greater, respectively). The community composition of the understory also varied by cross type, whereby additional understory plant species cluster with hybrids, not with parental species.
- *Conclusions*: Genetic composition dictated by hybridization in the overstory can play a role in structuring the associated understory plants in natural communities—where a hybridized overstory correlates with a species-rich understory—and thus can have cascading effects on community members of the same trophic level. The underlying mechanism requires further investigation.

Key words: canopy-understory interactions; community genetics; cottonwoods; hybridization; plant-plant interactions; *Populus angustifolia; Populus fremontii*; species-genetic diversity correlation.

Determining the factors that control the distribution and abundance of species is a central goal of ecology (e.g., Gilbert and Lechowicz, 2004). Understory plant species, particularly the herbaceous and grass communities, often represent a prominent component of the diversity of forests (Dirzo et al., 1992). Environmental characteristics such as light, water, and nutrients can vary at a fine spatial scale (i.e., on the order of meters), and these differing niches are known to influence the species composition of the understory (Tessier, 2007). Long recognized as important causes of these microhabitat differences in forests are not only the presence or absence of canopy trees (e.g., Price and Morgan, 2010) but also the particular species composition of the overstory (e.g., Ratliff et al., 1991; Knight et al., 2008).

otany

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doi:10.3732/ajb.1100137

For example, Ratliff et al. (1991) found that different overstory species in an oak woodland supported different herbaceous species under their canopies. It is not known, however, whether the manner in which understory plants in natural settings respond to genetic variation within closely related species and their hybrids is similar to how they respond to variation across species.

Hybridization is an important pathway of speciation for plants (Soltis and Soltis, 2009). Thus the ecological conditions that promote hybridization have been explored (e.g., Lepais et al., 2009), as have the effects of hybridization on the genetics of the plants themselves for several well-studied hybrid systems, including Eucalyptus (Potts and Dungey, 2004), mulberry (Morus; Burgess and Husband, 2006), and oak (Quercus; Lepais et al., 2009). However, the ecological effects of plant hybridization on other members of the community besides direct competitors (Vilà et al., 2000; Murrell et al., 2011) are much less understood and have been limited predominately to work on herbivorous insects and parasites (e.g., Whitham et al., 1994; Tovar-Sanchez and Oyama, 2006; Nakamura et al., 2010). In contrast, the *Populus* system has been a particularly fertile area for studying links between the genetic variation that results from hybridization with larger patterns of community structure and ecosystem processes (Whitham et al., 2006, 2008). Here we draw on ongoing work in this forest system to test whether the genetic identities of cottonwood trees influence understory plants in natural communities.

*Populus* spp. (Salicaceae, commonly known as *cottonwoods*, *aspens*, and *poplars*) are prevalent in riparian communities in the western United States and often function as foundational

American Journal of Botany 98(10): 1623–1632, 2011; http://www.amjbot.org/ © 2011 Botanical Society of America

<sup>&</sup>lt;sup>1</sup>Manuscript received 24 March 2011; revision accepted 29 July 2011.

The authors thank the Cottonwood Ecology Group, particularly J. Schweitzer, G. Wimp, L. Lamit, and J. Bailey, for valuable discussion and sharing information. B. King of the Utah Native Plant Society, J. Baker of the University of Utah's Red Butte Garden, and J. Arnold at Weber State University helped the authors identify and appreciate the understory vegetation of the Weber River watershed, and J. Brennan and D. Turner at Sanford University provided assistance with processing soil samples. This study was funded in part by a grant from the American Philosophical Society to R.I.A.

species, defined as "species that structure a community by creating locally stable conditions for other species, and by modulating and stabilizing fundamental ecosystem processes" (Dayton, 1972). Hybridization has been occurring in the *Populus* genus for at least 12 million years and happens naturally when the ranges of two or more *Populus* species overlap (Eckenwalder, 1984a, b). One such site of hybridization is along the Weber River in north-central Utah, USA, where Fremont cottonwoods (*Populus fremontii*), narrowleaf cottonwoods (*Populus angustifolia*), and their hybrids, which are together referred to as *cross types*, naturally occur.

The ecological effects of intraspecific genetic variation and hybridization in this system are beginning to be appreciated (Whitham et al., 2003). A suite of associated organisms, including terrestrial and aquatic arthropods (Wimp et al., 2004; LeRoy et al., 2006), birds (Martinsen and Whitham, 1994), mammals (Bailey et al., 2004), and soil microbes (Schweitzer et al., 2008), are affected by the identity of the different cross types. Moreover, hybridization in *Populus* can affect ecosystem processes, such as decomposition and nitrogen mineralization rates (Schweitzer et al., 2004, 2008) as well as water-use efficiency (Fischer et al., 2004), and can promote the evolution of new animal species living on these hybrid plants (Evans et al., 2008).

The microsite differences created by genetic variation within the foundational Populus spp. mosaic, to which diverse taxa occupying multiple trophic levels respond, also have the potential to influence within-trophic level interactions (e.g., Beatty, 1984). In a common garden study, Lamit et al. (2011) showed that hybridization between P. angustifolia and P. fremontii significantly affected understory biomass, species richness, and composition. Because common garden studies are designed to reduce environmental variation, it is important to determine whether geneticbased differences in common gardens can still be detected in the "wild," where environmental effects potentially could "swamp" genetic effects. We tested the prediction that the understory plant community in the wild also would respond to the genetic "footprint" of individual trees and, correspondingly, that the genetic effects would not be overshadowed by environmental effects. Responses from other members of the biotic community offer potential mechanisms by which understory plants also may be affected either directly or indirectly by the canopy tree, including variations in light (canopy architecture), water (water-use dynamics), and soil properties such as nutrient dynamics (decomposition and mineralization). If an effect exists, the nature of the relationship—whether hybrids elicit responses that are identical to, intermediate with, or lower than those of either of the parental species—would offer clues as to the likely mechanisms.

#### MATERIALS AND METHODS

Study system—Populus fremontii S. Wats (hereafter referred to as Fremont) dominates lower elevation riparian habitat in the western United States, and *P. angustifolia* James (hereafter *narrowleaf*) dominates higher-elevation habitat. Near Ogden, Utah (elevation 1300 m), the boundaries overlap, and both parental taxa are found along with their hybrids. We surveyed trees occurring within a 13-km stretch of the Weber River that is situated within the hybrid zone. The overstory canopy is dominated by *Populus* spp., representing nearly 70% of the individuals and over 90% of the biomass (Appendix 1). The understory, however, is more varied, composed of grasses, herbs, vines, saplings, and young trees, some of which are nonnative to the region.

*Genetic analyses*—All raw genetic data used in this study come from a study examining the effect of hybrid diversity on foliar arthropods (Wimp et al., 2004). For that study, Wimp et al. (2004) delimited 10 semidiscrete cottonwood

stands that are nearly identical in tree density, climate, and soil characteristics and are separated from each other by natural or human-made barriers (Schweitzer et al., 2011). Approximately 20 trees in each stand (each tree selected to be at least 30 m from another sampled tree) were genotyped at 48 amplified fragment length polymorphism (AFLP) loci (Wimp et al., 2004). For the current study, we calculated a hybrid index for each tree using those AFLP markers. The hybrid index defined the cross type of each tree and represents admixture coefficients generated under the admixture model in the program Structure (Pritchard et al., 2000). The number of subpopulations was limited to two because this study system contains two parental cottonwood species that readily hybridize. On the basis of this analysis, each tree received a hybrid index that represents the percentage of each individual composed of the two parental taxa. Hybrid index thus could be used to classify trees into discrete cross types (Fremont, narrowleaf, or hybrid) with 10% genetic admixture as the cut-off to delimit pure parental or hybrid status (Vähä and Primmer, 2006). That is, those trees with a hybrid index of <10% were classified as Fremonts, >90% as narrowleafs, and intermediate values as hybrids. We performed hybrid index calculations 10 times using an initial burn-in of 10<sup>4</sup> and 10<sup>6</sup> iterations. All runs then were combined with the CLUMPP program (Jakobsson and Rosenberg, 2007). Thus, although the raw data come from an earlier study (Wimp et al., 2004), the genetic analyses used to characterize parental taxa and their hybrids in the current study were generated for the current study.

**Plant surveys and edaphic sampling**—We surveyed under approximately 25 of each of the three cross types (Fremont, narrowleaf, and hybrids) across the 13-km stretch of the Weber River in July 2009. A  $0.5 \text{ m} \times 0.5 \text{ m}$  quadrat containing 25 intersection points was laid down 0.5 m from the base of the tree trunk at both the north and south cardinal directions. We recorded all vegetation touching a pin dropped at each intersection. Plants were identified by using *A Utah Flora* (Welsh, 2003), and the United States Department of Agriculture Natural Resources Conservation Service Plants Database (plants.usda.gov) was used as the final authority for nomenclature. Additionally, the depth of the leaf litter layer was measured at five points in the quadrat (the four corners plus the center), which was averaged per quadrat, and soil samples were collected by using a 15-cm-deep soil corer over a 2-day period. Botanical surveys and soil sampling were done in July 2009, and additional soil sampling occurred in August 2009.

Soil samples were frozen after collection and remained so until used for analysis. First, moisture was measured by drying 5 g of sieved (2 mm) soil in a 60°C drying oven for 48 hours. Soil moisture was calculated at two times during the growing season, and the mean was used for analysis. Second, total soil carbon and nitrogen were measured using a Carlo-Erba (Milan, Italy) 1500 CN Elemental Analyzer. Third, we determined soil pH by mixing soil with deionized water into a slurry and measuring it with a pH meter (Model 301, Orion Research Inc, Beverly, Massachusetts, USA.).

Statistical analysis—To test for spatial clustering of tree cross types, we performed an analysis of similarity (ANOSIM) on the Euclidean distances between trees by cross type within each stand. This was done using the "anosim" function of the vegan package (Oksanen et al., 2009) in the R programming language (R Development Core Team, 2009). A significant result indicates that within a stand, the trees of a particular cross type tend to be closer to each other than to trees of the other cross types.

Response variables of the understory plant community included species richness, Shannon diversity index, and cover. Shannon and Simpson diversity indices were correlated and show the same patterns; thus, for simplicity we use Shannon throughout. Cover was allowed to exceed 100% for a given species to avoid any spurious correlation between richness and cover. Understory diversity metrics were compared across cross types with ANOVA (after testing for assumptions of normality and homoscedasticity), and the location of tree (i.e., stand) was included as a block factor to account for any unmeasured variation across stands. The Weber River is a heavily invaded ecosystem, like many riparian communities in the United States (Stohlgren et al., 1998), and therefore vegetation was partitioned into native and exotic classes (Appendix 2). Furthermore, vegetation also was divided into the two dominant plant life forms, grasses and forbs (Appendix 2). Data were partitioned because, depending on the mechanism by which overstory interacts with understory, it is possible that origin status and growth form of the plant affect its response.

To test for differences in community composition among tree cross types, we used a permutation of multivariate analysis of variance test (PERMANOVA), which uses permutations to calculate significance of partitioning distance matrices among sources of variation (Anderson, 2001). This method was executed by using the "adonis" function of the vegan package in R with 1000 permutations and was implemented on both understory plant abundance and incidence (i.e., presence/absence) data. To get a representation of understory community composition, we used the nonmetric multidimensional scaling (NMDS) ordination technique with a matrix containing abundances of each understory species for each tree sampled. To graphically depict community structure as a 2-dimensional ordination, we used the "metaMDS" function with the Bray–Curtis dissimilarity coefficient in the R vegan package. Furthermore, indicator species analysis was used to assess whether individual species were associated with certain cross types. We used the method of Dufrene and Legendre (1997) as implemented in the "duleg" function of the labdsv R package (Roberts, 2007), also with 1000 iterations.

Lastly, to relate understory composition with genetic similarity, we followed the method of Bangert et al. (2006b), who found that genetic differences among individual cottonwoods affect the structure of arthropod communities. The genetic similarity rule asserts that the genetic identity of the plant influences the community structure of the dependent organisms, such that plants with a more similar genetic composition will support a more similar associated community (Bangert et al., 2006b). First, we calculated the Euclidean genetic distance between individual trees in the program GenAlEx (Peakall and Smouse, 2006) using the AFLP data generated by Wimp et al. (2004). This calculation produced a matrix of distances in which trees that are genetically similar have low values of distance. Secondly, we calculated a dissimilarity matrix of understory community using the Bray-Curtis distance metric, as implemented in the "vegdist" function of the vegan package in R. As with the genetic distance matrix, low values indicate compositional similarity. We then conducted a Mantel test on the two matrices to see whether any correlation existed between genetic distance of trees and compositional differences of the understory. The understory composition matrix was log-transformed before analysis.

## RESULTS

Overstory tree genetics—The vast majority of trees were readily classified as parentals or showed a near 50/50 split between the parental species and therefore most likely represented  $F_1$  hybrids. Although backcrossing of  $F_1$  hybrids to pure narrowleaf does occur (Martinsen et al., 2001), only one tree surveyed for understory vegetation showed an indication of backcrossing. Thus, the hybrid cross type class used in these analyses predominately represented F1 hybrids, though for clarity we refer to the "hybrid" cross type. The 48 AFLP markers showed all but two genotypes trees across the stands were unique individuals: two sampled F<sub>1</sub> hybrid trees in one stand exhibited the same genetic profile. The composition of the Populus taxa had not changed appreciably: only one tree that had been sampled for the 2004 study (Wimp et al., 2004) was found to be dead at the time of our understory sampling (2009), and this tree was not included in the analysis. Spatial analysis showed that within each stand, the tree cross types were not clustered (analysis of similarity for all stands,  $P \ge 0.09$ ).

**Understory plant community**—Vegetation surveys detected a rich understory of 67 species, composed of grasses, herbs, vines, and trees (Appendix 2). The Poaceae, Asteraceae, and Fabaceae were the most represented of the 30 families encountered. Approximately 40% of the identified species encountered are nonnative to the region (Appendix 2). Most of the species encountered were graminoid and forb species (83%), and the remainder were trees and shrubs. Over 80% of the identified species were perennial, and over 65% were dicots (Appendix 2).

Effects of hybridization on understory plant diversity metrics and edaphic properties—Cross type had a significant signal on the understory plant community (Table 1). Species richness, Shannon index, and cover were significantly higher under hybrids than under either parental species (Fig. 1). Under hybrids, plant species richness was 30-53% greater, plant diversity 40-48% greater, and plant cover 35-74% greater. A similar pattern was found when either exotic or native species in isolation were considered and when the two predominant growth forms (grasses and forbs) were considered in isolation (Fig. 1). The magnitude of the effect was greater on native than exotic plants (native: richness 46-69%, diversity 87-106%, and cover 117–151%; exotic: richness 27–49%, diversity 66–86%, and cover 0-32%) and on forbs compared with grasses (forbs: richness 43-98%, diversity 96-158%, cover 122%; grasses: richness: 30%, diversity: 48-58%, cover: 22-63%). Post hoc analyses revealed that in most cases, narrowleaf and Fremont understory diversity metrics were not significantly different from each other but were different from hybrids (Tukey-Kramer, P < 0.05, Fig. 1). For example, native vegetation does not differ under narrowleaf and Fremont cottonwoods for both diversity and cover, and both are significantly lower than hybrids. Stand location also had significant effects on understory diversity patterns (Table 1). This indicates that while the identity of the stand played a role in structuring understory plants communities, the differences across cross type generally emerged independent of the stand in which the tree was surveyed.

Depth of leaf litter (ANOVA,  $F_{2,60} = 0.770$ , P = 0.467), soil moisture (ANOVA,  $F_{2,60} = 1.147$ , P = 0.324), and soil C:N (ANOVA,  $F_{2,37} = 1.899$ , P = 0.164) did not vary by cross type. However, soil pH differed by cross type, with mean values 2–3% lower under hybrids than under parental species (ANOVA,  $F_{2,37} = 3.767$ , P = 0.032).

*Effects of hybridization on understory plant community composition*—Graphical representation of understory community

TABLE 1. Results of understory diversity metrics compared across *Populus* cross type (*P. fremontii*, *P. angustifolia*, and their hybrids) along the Weber River, Utah, USA, using ANOVA, with stand as a block factor. Significant values ( $\alpha < 0.5$ ) are in bold. Stand df = 9; cross type df = 2.

Diversity metric	All p	lants	Native plants		Non-native plants		Forbs		Grasses	
	F value	P value	F value	P value	F value	P value	F value	P value	F value	P value
Richness										
Stand	3.725	< 0.001	2.788	0.009	2.632	0.012	3.914	< 0.001	1.676	0.115
Cross type	4.971	0.010	2.586	0.084	2.805	0.069	1.408	0.253	2.983	0.058
Diversity										
Stand	4.024	< 0.001	1.929	0.065	3.753	< 0.001	4.483	< 0.001	2.811	0.008
Cross type	7.713	0.001	5.708	0.005	7.341	0.001	4.639	0.014	3.577	0.034
Cover										
Stand	3.447	0.002	2.809	0.008	2.155	0.038	3.781	< 0.001	3.346	0.002
Cross type	6.643	0.002	7.137	0.002	0.904	0.411	7.438	0.001	1.927	0.154



Fig. 1. Comparison of species richness, Shannon diversity index, and cover (%) reveal that hybrids of *Populus fremontii* (Fremont) and *P. angus-tifolia* (narrowleaf) along the Weber River in Utah, USA, have significantly greater diversity metrics of the understory plant community, compared with parental species. A significant effect of cross type when the factor of stands is blocked in the ANOVA is illustrated with an asterisk. Results are shown for all species pooled (three left panels), for exotic and native species in isolation, and for forbs and grasses in isolation (right panels). Data are means ( $\pm 1$  SE) grouped by cross type. Significant differences across cross type, as determined by post hoc analysis, are indicated with different letters.

composition demonstrated significant differences among cross type (Fig. 2). Moreover, statistical analysis showed that community composition was different under the broad category of cross types, after factoring out stand location, with both abundance (PERMANOVA  $F_{2,58} = 1.952$ , P = 0.016) and incidence data (PERMANOVA  $F_{2,69} = 2.956$ , P = 0.009). Pairwise comparisons of community composition indicated that the divergence between communities under hybrid and narrowleaf cottonwoods was driving this pattern (narrowleaf vs. hybrid, PERMANOVA  $F_{1,38} = 2.942$ , P = 0.004; Fremont vs. hybrid, PERMANOVA  $F_{1,31} = 1.087$ , P = 0.363; narrowleaf vs. Fremont, PERMANOVA  $F_{1,37} = 0.981$ , P = 0.444; Fig. 2).

Indicator species analysis highlighted five species as clustering under hybrids (4 native, 1 exotic) when all three cross types are considered as separate classes (Table 2A). Because compositional analysis indicated no difference under narrowleaf and Fremont parental trees, indicator species analysis also was done to compare understory composition under hybrids vs. under the pooled parental species. A further two species emerged as important under the hybrids (total 5 natives, 2 exotic; Table 2B). This contingent of species spanned several growth forms (including forbs, vines, grasses, and tree), life forms (annuals and perennials), and origin (native/exotic) status.

Finally, our findings support the genetic similarity rule (Bangert et al., 2006a, b; Barbour et al., 2009), in which genetically similar trees support more similar understory communities than do genetically dissimilar trees. Across cross types, the understory composition tended to diverge as the genetic distance between individual trees increased (Mantel *r* value = 0.076, P = 0.008). However, testing the genetic similarity rule among trees within each of the cross types showed no correlation (narrowleaf only: Mantel *r* value = 0.064, P = 0.298; Fremont only: Mantel *r* value = -0.012, P = 0.515).

#### DISCUSSION

Overstory and understory linkage-We found evidence to support our hypothesis that understory community composition and diversity were influenced by the genetics of the overstory foundation cottonwoods in wild settings. First, richness, diversity, and cover of understory plants were 30-74% higher under the hybrids between *P. fremontii* and *P. angustifolia* than under the parental species. Second, understory plant community composition clustered by overstory cross type, and the community under hybrids represented a unique assemblage rather than a merging of the parental communities. The distribution of understory vegetation in natural settings is obviously complex and dynamic (e.g., Tessier, 2007), and by demonstrating a consistent and significant linkage between canopy tree cross type and the herbaceous community, we show that the genetic differences in the overstory that result from hybridization should be included as an influential factor.

Although unmeasured environmental factors as causative agents cannot be ruled out in this observational study in the wild, we demonstrate a clear link between overstory genetic composition and understory species diversity. Several findings argue that this link is influenced by the genetic make-up of the tree. (1) The environment conditions along the brief stretch of the Weber River where the trees were sampled are remarkably similar (Schweitzer et al., 2011), and thus abiotic differences among stands and trees should be minimal. Indeed, the signal of tree identity on understory appeared independent of stand identity. Moreover, processes known to influence diversity at regional scales, such as dispersal limitation, are unlikely to operate in the fine scale of this geographic area. (2) All trees fall within the recognized hybrid zone and demonstrate no obvious spatial structure. Hybrids are spatially intermixed with Fremont and narrowleaf cottonwoods, appear in the center and periphery of stands, and are situated proximate and farther from the river, etc. Also, this hybrid zone is naturally occurring, formed at mid elevation where ranges overlap and not from an area of high disturbance. (3) Although it is possible that extremely localized patches of disturbance could produce a correlation between overstory genetics and understory species, we find this unlikely given the lack of spatial structure in cross types: the highly localized (i.e., on the order of meters) disturbance would have to be persistent across years, such that the disturbance that favored the establishment of hybrids in a particular location in the past continues to influence understory plants today. Such parallel effects of disturbance on species diversity of plants and hybridization between plants in one locality would be predicted (Petraitis et al., 1989; McKinnon, 2005) though to our knowledge has never been documented. Experimental manipulation of understory plants in natural stands would explicitly test the hypothesis that it is the tree cross type that shapes the understory plant community.

Plant-plant interactions-A positive association between overstory diversity and the associated organisms lends support for the "diversity begets diversity" hypothesis of Whittaker (1975), and several genetic and demographic processes in tree hybrid zones could affect understory community structure in specific ways. As was hypothesized for the effect of hybridization on phytophagous insects (Whitham et al., 1994), interbreeding could break down genetically determined chemical composition and consequently disrupt any negative interactions with community members. Similarly, because hybrid zones are often on the periphery of the parent species ranges, parental species may be physiologically stressed and thus less effective at enforcing negative or positive interactions. From a genetic perspective, introgression between parental taxa is often asymmetrical because of factors such as differences in species abundances (Burgess et al., 2005) or asynchronous phenology (Belahbib et al., 2001). If one of the parental species structures the understory community in a more dominant manner, then the understory could be "pushed" to match the genetically overriding taxa. Moreover, hybrid zones can support the maintenance and evolution of interacting species that associate specifically

with hybrids (Fritz et al., 1994) and thus represent areas of special concern from a conservation perspective.

Considering beyond the diversity that results from hybridization specifically, the field of community genetics (Agrawal, 2003) has increasingly highlighted the importance that genetic identity and genetic diversity within species and species complexes have for within-trophic interactions in a community ecology context. For instance, phenotypic differences within an overstory tree and a shrub affect understory plants (Pakeman et al., 2006; Crutsinger et al., 2010), as does variation within dominant herbaceous plants (Lankau and Strauss, 2007; Crutsinger et al., 2008; Michalet et al., 2011). The causes of this influence differ by system and include variation in the chemical diversity of monoterpenes (Iason et al., 2005; Pakeman et al., 2006) and canopy architecture (Crutsinger et al., 2010). Thus, differences between plant taxa-ranging from different species, to closely related species and their hybrids, and even within individuals at the population level-are associated with microsite differences to which the understory species demonstrate a niche-based response.

**Potential mechanisms**—Light, water, and soil properties are the predominant factors generally considered to determine understory plant community diversity metrics (e.g., Dirzo et al., 1992) and to limit understory plant growth (e.g., Shirley, 1945), and here we discuss each of these in turn for the *Populus* system. Architecture does vary among the cross types, where narrowleaf canopies are tapered and Fremont canopies wide (Martinsen and Whitham, 1994). In a common garden, it was shown that photosynthetically active radiation levels decrease with hybrid index (i.e., lower light under Fremont); however, this variation in light did not correlate with understory plant richness or composition (Lamit et al., 2011). Indeed, if light is



Axis 1

Fig. 2. Plot of understory plant community composition under *Populus fremontii* (Fremont), *P. angustifolia* (narrowleaf), and their hybrids along the Weber River in Utah, USA, based on nonmetric multidimensional scaling. Represented are the mean and standard errors of the ordination axis scores for the three categories, showing differences in the understory plant community composition under each tree cross type.

the dominant factor in determining understory plants in natural *Populus* stands, then tree-level surveys would have shown narrowleaf to have the highest understory species diversity. The absence of this pattern in the wild strongly suggests that light is not the prominent mechanism, possibly because individual tree canopies may not consistently vary or because larger canopy dynamics supersede the influence.

Second, belowground processes also are important to consider. Plant secondary compound concentrations, particularly condensed tannins, are lower in the leaves of Fremont than narrowleaf cottonwoods, and condensed tannins are strongly tied to litter decomposition and nitrogen mineralization rates (Schweitzer et al., 2004). Furthermore, in apparent response to these effects of litter quality on decomposition and nutrient cycling, Fischer et al. (2006) found that fine root production varied greatly among cross types in which narrowleaf cottonwoods invested the most, Fremont the least, and hybrids were intermediate. Thus, one could predict that genotype-determined moderation of the nutrient pool available to plants and competition with fine roots via condensed tannins could alter understory communities. In this case, the pattern observed would again be linear, with hybrids intermediate between the parents. This linear pattern is the observation made in a common garden study, in which the authors demonstrate a correlative link between foliar condensed tannins in Populus and the understory (Lamit et al., 2011). However, we did not observe a linear pattern in the natural stands, and the discrepancy in the response of understory plants to overstory genetic identity by location-either a common garden or natural stands-indicates that although understory species and tree genetic diversity correspond in nonrandom ways, the particular mechanism for the correlation may be context dependent. For instance, in the common garden, Populus trees are spaced at regular intervals, and it is possible that under uniform conditions, hybrids can modulate belowground processes via leaf litter; in the wild, however, variable spacing of trees could respond to or, more likely, contribute to a highly patchy environment that overwhelms differences due to variation in condensed tannin production. In addition to differences in plant secondary compounds, soil pH levels can be associated with different tree species (Crozier and Boerner, 1984), and here we find that soils underneath hybrids are more acidic. Plant-induced changes in soil pH can occur through a variety of mechanisms, including changes in nitrification rate

and cation redistribution (Ehrenfeld et al., 2001). The mechanistic link between cross type and soil pH in this particular system remains unclear.

Third, water remains an intriguing possibility, as hybrids are known to exhibit higher water use than either parental types (Fischer et al., 2004). Greater water use is expected to be associated with a greater hydraulic lift (also known as hydraulic redistribution), the process by which deep-rooted plants bring water from lower soil layers to upper soil layers (Horton and Hart, 1998). It is possible that hydraulic lift may still be occurring—and producing the observed outcome by increasing soil moisture under hybrids relative to parental taxa-but the increased water availability to neighboring plants was undetectable with our sampling methodology. Hydraulic lift, too, is a potential mechanism that could differ between wild and common garden settings: in the common garden, the regular and relatively closely situated trees (5-7 m; Lamit et al., 2011) could negate the possible influence of soil moisture on understory plants; whereas in the wild, differences in hydraulic lift among patchily distributed trees could demonstrate a marked influence on understory plants.

Several facets of the analysis do offer support for potential mechanisms behind the link between overstory and understory in natural settings. First, the effect of hybridization on understory is not linear (where hybrids are intermediate between parental species). Instead, this signal for heterosis could imply a breakdown of process that otherwise limits understory growth, such as that which occurred for the generation of allelopathic compounds in Scots pine trees (Iason et al., 2005), or the strengthening of process that fosters understory growth, such as hydraulic lift. Second, the fact that the pattern is consistently followed for both native and exotic plants, as well as for variable life forms (e.g., forbs, grasses), is evidence that the effect is rather general. Therefore, we would expect the mechanism to draw on a common, nonspecific strategy for plant growth. Third, indicator species analysis revealed that many understory species are associated with the hybrids specifically, indicating that the species that define composition under the hybrids are not a random sample of the environment under the parentals but rather contain a unique assemblage of additional plant species. Moreover, it is not simply that community composition differs by cross type but also by richness, diversity, and cover (unlike foliar arthropods; Wimp et al., 2004, 2005). This indicates that the area under hybrids is overall more hospitable to plant growth. Fourth, little indication exists for a role

Table 2.	Indicator	species	analysis	comparin	ng comm	unity	composition	(A) an	nong th	e three	Populus	cross	types	and (E	<ol><li>B) between</li></ol>	hybrids	and the
parent	tal species	. IV = in	ndicator v	alue (corr	respondin	g P va	alue is shown	); clust	er = the	group	with which	ch the	unders	tory sp	becies is as	sociated.	

Understory Species	IV	Р	Cluster	Status	Growth form	Duration
(A) Three groupings: Narrowleaf–hybrid–Fremont						
Ambrosia artemisiifolia	0.506	0.001	Hybrid	Native	Forb, herb	Annual
Clematis ligusticifolia	0.158	0.025	Hybrid	Native	Vine	Perennial
Elymus glaucus	0.462	0.002	Hybrid	Native	Grass	Perennial
Lactuca serriola	0.242	0.026	Hybrid	Exotic	Forb, herb	Annual, biennial
Pseudoroegneria spicata	0.182	0.014	Hybrid	Native	Grass	Perennial
(B) Two groupings: Hybrid-parental taxa						
Ambrosia artemisiifolia	0.527	0.001	Hybrid	Native	Forb/Herb	Annual
Clematis ligusticifolia	0.171	0.008	Hybrid	Native	Vine	Perennial
Elymus glaucus	0.537	0.001	Hybrid	Native	Grass	Perennial
Lactuca serriola	0.258	0.018	Hybrid	Exotic	Forb, herb	Annual, biennial
Melilotus officinalis	0.134	0.134	Hybrid	Exotic	Forb, herb	Annual, biennial, perennial
Pseudoroegneria spicata	0.182	0.014	Hybrid	Native	Grass	Perennial
Rosa woodsii	0.131	0.131	Hybrid	Native	Subshrub	Perennial

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of other biotic factors influencing particular understory species. Although more bird nests appear in hybrids (Martinsen and Whitham, 1994), few of the indicator species associated with hybrids are bird dispersed. In total, these factors indicate that the trees seem to alter the microenvironment in a broad way to which understory plants in general respond, and soil-based properties, particularly increased water availability brought about through hydraulic lift or changes in soil pH, remain promising hypotheses to test.

**Conclusions**—We have surveyed understory vegetation in a natural riparian community to show an association between tree cross type within hybridizing *Populus* and the associated understory plant community. Hybrid trees are characterized by a more diverse understory and a unique assemblage of understory plants compared with either of the parental tree species. We show that understory plant communities can consistently respond to microsite habitat differences associated with particular tree cross types, and the precise underlying mechanism represents a promising line of further study.

## LITERATURE CITED

- AGRAWAL, A. A. 2003. Community genetics: New insights into community ecology by integrating population genetics. *Ecology* 84: 543–544.
- ANDERSON, M. J. 2001. A new method for non-parametric multivariate analysis of variance. Austral Ecology 26: 32–46.
- BAILEY, J. K., J. A. SCHWEITZER, B. J. REHILL, R. L. LINDROTH, G. D. MARTINSEN, AND T. G. WHITHAM. 2004. Beavers as molecular geneticists: A genetic basis to the foraging of an ecosystem engineer. *Ecology* 85: 603–608.
- BANGERT, R. K., G. J. ALLAN, R. J. TUREK, G. M. WIMP, N. MENESES, G. D. MARTINSEN, P. KEIM, ET AL. 2006a. From genes to geography: A genetic similarity rule for arthropod community structure at multiple geographic scales. *Molecular Ecology* 15: 4215–4228.
- BANGERT, R. K., R. J. TUREK, B. J. REHILL, G. M. WIMP, J. A. SCHWEITZER, G. J. ALLAN, J. K. BAILEY, ET AL. 2006b. A genetic similarity rule determines arthropod community structure. *Molecular Ecology* 15: 1379–1391.
- BARBOUR, R. C., J. M. O'REILLY-WAPSTRA, D. W. DE LITTLE, G. J. JORDAN, D. A. STEANE, J. R. HUMPHREYS, J. K. BAILEY, ET AL. 2009. A geographic mosaic of genetic variation within a foundation tree species and its community-level consequences. *Ecology* 90: 1762–1772.
- BEATTY, S. W. 1984. Influence of microtopography and canopy species on spatial patterns of forest understory plants. *Ecology* 65: 1406–1419.
- BELAHBIB, N., M. H. PEMONGE, A. OUASSOU, H. SBAY, A. KREMER, AND R. J. PETIT. 2001. Frequent cytoplasmic exchanges between oak species that are not closely related: *Quercus suber* and *Q. ilex* in Morocco. *Molecular Ecology* 10: 2003–2012.
- BURGESS, K. S., AND B. C. HUSBAND. 2006. Habitat differentiation and the ecological costs of hybridization: The effects of introduced mulberry (*Morus alba*) on a native congener (*M. rubra*). *Journal of Ecology* 94: 1061–1069.
- BURGESS, K. S., M. MORGAN, L. DEVERNO, AND B. C. HUSBAND. 2005. Asymmetrical introgression between two *Morus* species (*M. alba, M. rubra*) that differ in abundance. *Molecular Ecology* 14: 3471–3483.
- CROZIER, C. R., AND R. E. J. BOERNER. 1984. Correlations of understory herb distribution patterns with microhabitats under different tree species in a mixed mesophytic forest. *Oecologia* 62: 337–343.
- CRUTSINGER, G. M., L. SOUZA, AND N. J. SANDERS. 2008. Intraspecific diversity and dominant genotypes resist plant invasions. *Ecology Letters* 11: 16–23.
- CRUTSINGER, G. M., S. Y. STRAUSS, AND J. A. RUDGERS. 2010. Genetic variation within a dominant shrub species determines plant species colonization in a coastal dune ecosystem. *Ecology* 91: 1237–1243.

- DAYTON, P. K. 1972. Toward an understanding of community resilience and the potential effects of enrichments to the benthos at McMurdo Sound, Antarctica. *In* B. C. Parker [ed.], Proceedings of the colloquium on conservation problems in Antarctica, 81–96. Allen Press, Lawrence, Kansas, USA.
- DIRZO, R. D., C. C. HORVITZ, H. QUEVEDO, AND M. A. LÓPEZ. 1992. The effects of gap size and age on the understorey herb community of a tropical Mexican rain forest. *Journal of Ecology* 80: 809–822.
- DUFRENE, M., AND P. LEGENDRE. 1997. Species assemblages and indicator species: The need for a flexible asymmetrical approach. *Ecological Monographs* 67: 345–366.
- ECKENWALDER, J. E. 1984a. Natural intersectional hybridization between North American species of *Populus* (Salicaceae) in sections Aigeiros and Tacamahaca. II. Taxonomy. *Canadian Journal of Botany* 62: 325–335.
- ECKENWALDER, J. E. 1984b. Natural intersectional hybridization between North American species of *Populus* (Salicaceae) in sections Aigeiros and Tacamahaca. III. Paleobotany and evolution. *Canadian Journal* of *Botany* 62: 336–342.
- EHRENFELD, J. G., R. KOURTEV, AND W. HUANG. 2001. Changes in soil functions following invasions of exotic understory plants in deciduous forests. *Ecological Applications* 11: 1287–1300.
- EVANS, L. M., G. J. ALLAN, S. M. SHUSTER, S. A. WOOLBRIGHT, AND T. G. WHITHAM. 2008. Tree hybridization and genotypic variation drive cryptic speciation of a specialist mite herbivore. *Evolution*; *International Journal of Organic Evolution* 62: 3027–3040.
- FISCHER, D., S. C. HART, B. J. REHILL, R. L. LINDROTH, AND T. G. WHITHAM. 2006. Do high-tannin leaves require more roots? *Oecologia* 149: 668–675.
- FISCHER, D. G., S. C. HART, T. G. WHITHAM, G. D. MARTINSEN, AND P. KELM. 2004. Ecosystem implications of genetic variation in wateruse of a dominant riparian tree. *Oecologia* 139: 288–297.
- FRITZ, R. S., C. M. NICHOLSORIANS, AND S. J. BRUNSFELD. 1994. Interspecific hybridization of plants and resistance to herbivores: Hypotheses, genetics, and variable responses in a diverse herbivore community. *Oecologia* 97: 106–117.
- GILBERT, B., AND M. J. LECHOWICZ. 2004. Neutrality, niches, and dispersal in a temperate forest understory. *Proceedings of the National Academy of Sciences, USA* 101: 7651–7656.
- HORTON, J. L., AND S. C. HART. 1998. Hydraulic lift: A potentially important ecosystem process. *Trends in Ecology & Evolution* 13: 232–235.
- IASON, G. R., J. J. LENNON, R. J. PAKEMAN, V. THOSS, J. K. BEATON, D. A. SIM, AND D. A. ELSTON. 2005. Does chemical composition of individual Scots pine trees determine the biodiversity of their associated ground vegetation? *Ecology Letters* 8: 364–369.
- JAKOBSSON, M., AND N. A. ROSENBERG. 2007. CLUMPP: A cluster matching and permutation program for dealing with label switching and multimodality in analysis of population structure. *Bioinformatics* (Oxford, England) 23: 1801–1806.
- KNIGHT, K. S., J. OLEKSYN, A. M. JAGODZINSKI, P. B. REICH, AND M. KASPROWICZ. 2008. Overstorey tree species regulate colonization by native and exotic plants: A source of positive relationships between understorey diversity and invasibility. *Diversity and Distributions* 14: 666–675.
- LAMIT, L. J., T. WOJTOWICZ, Z. KOVACS, S. C. WOOLEY, M. ZINKGRAF, T. G. WHITHAM, R. L. LINDROTH, ET AL. 2011. Hybridization among foundation tree species influences the structure of associated understory plant community. *Botany* 89: 165–174.
- LANKAU, R. A., AND S. Y. STRAUSS. 2007. Mutual feedbacks maintain both genetic and species diversity in a plant community. *Science* 317: 1561–1563.
- LEPAIS, O., R. J. PETIT, E. GUICHOUX, J. E. LAVABRE, F. ALBERTO, A. KREMER, AND S. GERBER. 2009. Species relative abundance and direction of introgression in oaks. *Molecular Ecology* 18: 2228–2242.
- LEROY, C. J., T. G. WHITHAM, P. KEIM, AND J. C. MARKS. 2006. Plant genes link forest and streams. *Ecology* 87: 255–261.
- MARTINSEN, G. D., AND T. G. WHITHAM. 1994. More birds nest in hybrid cottonwood trees. *Wilson Bulletin* 106: 474–481.

- MARTINSEN, G. D., T. G. WHITHAM, R. J. TUREK, AND P. KEIM. 2001. Hybrid populations selectively filter gene introgression between species. *Evolution; International Journal of Organic Evolution* 55: 1325–1335.
- MCKINNON, G. 2005. Reticulate evolution in higher plants. In R. J. Henry [ed.], Plant diversity and evolution: Genotypic and phenotypic variation in higher plants, 81–96. CABI Publishing, Cambridge, Massachusetts, USA.
- MICHALET, R., S. XIAO, B. TOUZARD, D. S. SMITH, L. A. CAVIERES, R. M. CALLAWAY, AND T. G. WHITHAM. 2011. Phenotypic variation in nurse traits and community feedbacks define an alpine community. *Ecology Letters* 14: 433–443.
- MURRELL, C., E. GERBER, C. KREBS, M. PAREPA, U. SCHAFFNER, AND O. BOSSDORF. 2011. Invasive knotweed affects native plants through allelopathy. *American Journal of Botany* 98: 38–43.
- NAKAMURA, M., M. ASANUMA, AND T. HIURA. 2010. Differential effects of host plant hybridization on herbivore community structure and grazing pressure on forest canopies. *Oikos* 119: 1445–1452.
- OKSANEN, J., R. KINDT, P. LEGENDRE, B. O'HARA, G. L. SIMPOSON, P. SOLYMOS, M. H. H. STEVENS, AND H. WAGNER. 2009. vegan: Community Ecology Package. R package version 1.15-4.
- PAKEMAN, R. J., J. K. BEATON, V. THOSS, J. J. LENNON, C. D. CAMPBELL, D. WHITE, AND G. R. IASON. 2006. The extended phenotype of Scots pine *Pinus sylvestris* structures the understorey assemblage. *Ecography* 29: 451–457.
- PEAKALL, R., AND P. E. SMOUSE. 2006. GENALEX 6: Genetic analysis in Excel. Population genetic software for teaching and research. *Molecular Ecology Notes* 6: 288–295.
- PETRAITIS, P. S., R. E. LATHAM, AND R. A. NIESENBAUM. 1989. The maintenance of species diversity by disturbance. *Quarterly Review of Biology* 64: 393–418.
- POTTS, B. M., AND H. S. DUNGEY. 2004. Interspecific hybridization of *Eucalyptus*: Key issues for breeders and geneticists. *New Forests* 27: 115–138.
- PRICE, J. N., AND J. W. MORGAN. 2010. Small-scale patterns of species richness and floristic composition in relation to microsite variation in herb-rich woodlands. *Australian Journal of Botany* 58: 271–279.
- PRITCHARD, J. K., M. STEPHENS, AND P. DONNELLY. 2000. Inference of population structure using multilocus genotype data. *Genetics* 155: 945–959.
- R DEVELOPMENT CORE TEAM. 2009. R: A language and environment for statistical computing. R Foundation for Statistical Computing, Vienna, Austria.
- RATLIFF, R. D., D. A. DUNCAN, AND S. E. WESTFALL. 1991. California oak-woodland overstory species affect herbage understory: Management implications. *Journal of Range Management* 44: 306–310.
- ROBERTS, D. W. 2007. labdsv: Ordination and Multivariate Analysis for Ecology. R package version 1.3-1.
- SCHWEITZER, J. A., J. K. BAILEY, B. J. REHILL, G. D. MARTINSEN, S. C. HART, R. L. LINDROTH, P. KEIM, ET AL. 2004. Genetically based trait in a dominant tree affects ecosystem processes. *Ecology Letters* 7: 127–134.
- SCHWEITZER, J. A., D. G. FISCHER, B. J. REHILL, S. C. WOOLEY, S. WOOLBRIGHT, R. L. LINDROTH, T. G. WHITHAM, ET AL. 2011. Forest gene diversity is correlated with the composition and function of soil microbial communities. *Population Ecology* 53: 35–46.

- SCHWEITZER, J. A., M. D. MADRITCH, J. K. BAILEY, C. J. LEROY, D. G. FISCHER, B. J. REHILL, R. L. LINDROTH, ET AL. 2008. From genes to ecosystems: The genetic basis of condensed tannins and their role in nutrient regulation in a *Populus* model system. *Ecosystems (New York, N.Y.)* 11: 1005–1020.
- SHIRLEY, H. L. 1945. Reproduction of upland conifers in the Lake states as affected by root competition and light. *American Midland Naturalist* 33: 537–612.
- SOLTIS, P. S., AND D. E. SOLTIS. 2009. The role of hybridization in plant speciation. *Annual Review of Plant Biology* 60: 561–588.
- STOHLGREN, T. J., K. A. BULL, Y. OTSUKI, C. A. VILLA, AND M. LEE. 1998. Riparian zones as havens for exotic plant species in the central grasslands. *Plant Ecology* 138: 113–125.
- TESSIER, J. T. 2007. Re-establishment of three dominant herbaceous understory species following fine-scale disturbance in a Catskill northern hardwood forest. *Journal of the Torrey Botanical Society* 134: 34–44.
- TOVAR-SANCHEZ, E., AND K. OYAMA. 2006. Community structure of canopy arthropods associated to *Quercus crassifolia* × *Quercus crassipes* complex. *Oikos* 112: 370–381.
- VÄHÄ, J. P., AND C. R. PRIMMER. 2006. Efficiency of model-based Bayesian methods for detecting hybrid individuals under different hybridization scenarios and with different numbers of loci. *Molecular Ecology* 15: 63–72.
- VILÀ, M., E. WEBER, AND C. M. D'ANTONIO. 2000. Conservation implications of invasion by plant hybridization. *Biological Invasions* 2: 207–217.
- WELSH, S. L. 2003. A Utah flora, 3rd ed. Brigham Young University, Provo, Utah, USA.
- WHITHAM, T. G., J. K. BAILEY, J. A. SCHWEITZER, S. M. SHUSTER, R. K. BANGERT, C. J. LEROY, E. V. LONSDORF, ET AL. 2006. A framework for community and ecosystem genetics: From genes to ecosystems. *Nature Reviews Genetics* 7: 510–523.
- WHITHAM, T. G., S. P. DIFAZIO, J. A. SCHWEITZER, S. M. SHUSTER, G. J. ALLAN, J. K. BAILEY, AND S. A. WOOLBRIGHT. 2008. Perspective— Extending genomics to natural communities and ecosystems. *Science* 320: 492–495.
- WHITHAM, T. G., P. A. MORROW, AND B. M. POTTS. 1994. Plant hybrid zones as centers of biodiversity—The herbivore community of 2 endemic Tasmanian eucalypts. *Oecologia* 97: 481–490.
- WHITHAM, T. G., W. P. YOUNG, G. D. MARTINSEN, C. A. GEHRING, J. A. SCHWEITZER, S. M. SHUSTER, G. M. WIMP, ET AL. 2003. Community and ecosystem genetics: A consequence of the extended phenotype. *Ecology* 84: 559–573.
- WHITTAKER, R. H. 1975. Communities and ecosystems. MacMillan Publishing Company, New York, New York, USA.
- WIMP, G. M., G. D. MARTINSEN, K. D. FLOATE, R. K. BANGERT, AND T. G. WHITHAM. 2005. Plant genetic determinants of arthropod community structure and diversity. *Evolution; International Journal of Organic Evolution* 59: 61–69.
- WIMP, G. M., W. P. YOUNG, S. A. WOOLBRIGHT, G. D. MARTINSEN, P. KEIM, AND T. G. WHITHAM. 2004. Conserving plant genetic diversity for dependent animal communities. *Ecology Letters* 7: 776–780.

APPENDIX 1. At 10 cottonwood stands in the *Populus fremontii* × angustifolia hybrid zone, woody vegetation encountered along six 40-m transects was recorded to assess overstory composition. *Populus* is uniformly dominant in number and biomass.

					Importance						Importance
Stand	Species	Density <sup>a</sup>	Frequency <sup>b</sup>	Dominance <sup>c</sup>	value <sup>d</sup>	Stand	Species	Density <sup>a</sup>	Frequency <sup>b</sup>	Dominance <sup>c</sup>	value <sup>d</sup>
ESH	Acer negundo	4	0.50	1790.23	41.06		Populus spp.	25	1.00	14140.50	204.61
1 511	Crataeous monoovna	1	0.17	41 59	10.94		Rhus trilobata	1	0.17	25.16	9.20
	Populus snn	17	0.83	30504.92	184 97		Rosa woodsii	2	0.33	25.67	18.23
	Prunus viroiniana	1	0.03	41 59	10.94		Salix spp.	3	0.33	124.76	21.69
	Rosa woodsii	3	0.33	38 51	24.88	H6	Elaeagnus angustifolia	3	0.33	124.76	24.36
	Salix spp.	6	0.17	249.51	27.21		Malus spp.	1	0.17	25.16	10.60
H1	Acer negundo	3	0.50	218.20	23.46		Populus spp.	25	1.00	18409.95	210.49
	Cornus sericea	2	0.33	50.31	14 94		Salix spp.	6	0.50	827.60	44.01
	Crataeous monoovna	3	0.50	124 76	22.78		Tamarix chinensis	1	0.17	12.84	10.54
	Elaeaonus anoustifolia	1	0.17	41 59	7 59	NU1	Acer negundo	1	0.17	51.34	8.00
	Mahonia renens	1	0.17	25.16	7 47		Elaeagnus angustifolia	1	0.17	10.40	7.75
	Populus spp	16	1.00	13082.05	168.42		Juniper chinensis	1	0.17	41.59	7.94
	Prunus viroiniana	10	0.17	25.16	7 47		Populus spp.	33	1.00	15685.20	200.39
	Rhus trilohata	6	0.83	150.94	40.49		Prunus americana	2	0.33	83.17	15.88
	Rosa woodsii	1	0.03	12.84	7 38		Rhus trilobata	3	0.33	75.47	17.96
нз	Acer negundo	8	0.67	179.43	30.17		Salix spp.	6	0.83	249.51	42.08
115	Cornus sericea	4	0.50	52.88	18 78	NU2	Juniper chinensis	1	0.17	41.59	10.96
	Flagganus angustifolia	1	0.17	86.76	5.97		Populus spp.	9	0.83	3616.40	146.91
	Fravinus nennsylvanica	2	0.33	66.87	11.48		Quercus gambelii	2	0.33	166.86	23.86
	Iuniner chinensis	1	0.17	37.09	5 76		Rhus trilobata	15	0.83	377.35	92.05
	Juniperus communis	1	0.17	0.51	5.60		Salix spp.	3	0.33	124.76	26.22
	Mahonia renens	1	0.17	25.16	5.00	S9	Acer negundo	1	0.17	51.34	7.55
	Populus spp	25	1.00	22917.62	163 29		Crataegus monogyna	12	0.33	184.95	30.30
	Rhus trilohata	25	0.17	37.00	7 51		Elaeagnus angustifolia	3	0.33	786.02	18.91
	Rosa woodsii	3	0.33	38 51	13.12		Populus spp.	40	1.00	27688.54	189.09
	Salir spp	0	0.55	340.90	32.61		Rhus trilobata	2	0.33	50.31	14.92
Н/	Flacanus angustifolia	5	0.50	281.86	28.26		Rosa woodsii	2	0.33	25.67	14.84
114	Fravinus pannsylvanica	2	0.17	83.17	0.20		Salix spp.	6	0.17	249.51	15.69
	Populus spp	42	1.00	5020 70	9.71 187.78		Ulmus pumila	1	0.17	391.47	8.71
	Salix spp.	16	1.00	02.80	56 70	SU	Acer negundo	1	0.17	12.84	9.20
	Tamarix chinansis	10	0.50	30.20	20.75		Crataegus monogyna	1	0.17	41.59	9.32
Ц5	Acar nagundo	1	0.50	51.34	0.38		Elaeagnus angustifolia	2	0.33	83.17	18.65
115	Corrus saricaa	2	0.17	28.24	18 25		Fraxinus pennsylvanica	1	0.17	51.34	9.37
	Elacaphus angustifalia	ے 1	0.33	20.24	0.20		Populus spp.	41	1.00	22687.88	223.47
	Lucignus angustijolla	1	0.17	41.59	9.31		Quercus gambelii	1	0.17	4.62	9.16
	jumper chinensis	1	0.17	41.59	9.31		Salix spp.	3	0.33	124.76	20.83

<sup>a</sup>Number of individuals.

<sup>b</sup>Proportion of transects in which taxon appears.

<sup>c</sup> Sum of basal area (cm) for all individuals of a taxon.

<sup>d</sup>Sum of all three relativized values. Relativized values are the percentage of the metric (density, frequency, dominance) that each taxa represents in each stand and thus sums to 300 for each stand.

APPENDIX 2. Table of plant species composing understory community in the *Populus fremontii* and *P. angustifolia* hybrid zone along the Weber River in north-central Utah, USA.

Species	Family	Native or exotic	Growth form	Duration	Group
Acer negundo	Sapindaceae	Native	Tree	Perennial	Dicot
Achillea millefolium	Asteraceae	Native	Forb	Perennial	Dicot
Aegilops cylindrica	Poaceae	Exotic	Grass	Annual	Monocot
Agastache urticifolia	Lamiaceae	Native	Forb	Perennial	Dicot
Agropyron cristatum	Poaceae	Exotic	Grass	Perennial	Monocot
Ambrosia artemisiifolia	Asteraceae	Native	Forb	Annual	Dicot
Apocynum cannabinum	Apocynaceae	Native	Forb	Perennial	Dicot
Artemisia ludoviciana	Asteraceae	Native	Forb	Perennial	Dicot
Artemisia tridentata	Asteraceae	Native	Shrub, tree	Perennial	Dicot
Asclepias speciosa	Asclepiadaceae	Native	Forb	Perennial	Dicot
Astragalus utahensis	Fabaceae	Native	Forb	Perennial	Dicot
Bromus carinatus	Poaceae	Native	Grass	Annual, perennial	Monocot
Bromus diandrus	Poaceae	Exotic	Grass	Annual, perennial	Monocot
Bromus erectus	Poaceae	Exotic	Grass	Perennial	Monocot
Bromus inermis	Poaceae	Exotic	Grass	Perennial	Monocot
Bromus tectorum	Poaceae	Exotic	Grass	Annual	Monocot
Cardaria draba	Brassicaceae	Exotic	Forb	Perennial	Dicot

## APPENDIX 2. Continued.

Species	Family	Native or exotic	Growth form	Duration	Group
Cirsium arvense	Asteraceae	Exotic	Forb	Perennial	Dicot
Clematis ligusticifolia	Ranunculaceae	Native	Vine	Perennial	Dicot
Convolvulus arvensis	Convolvulaceae	Exotic	Vine, forb	Perennial	Dicot
Crataegus monogyna	Rosaceae	Exotic	Shrub, tree	Perennial	Dicot
Cynoglossum officinale	Boraginaceae	Exotic	Forb	Biennial	Dicot
Dactylis glomerata	Poaceae	Exotic	Grass	Perennial	Monocot
Digitaria sanguinalis	Poaceae	Native	Grass	Annual	Monocot
Dipsacus fullonum	Dipsacaceae	Exotic	Forb	Biennial	Dicot
Eleocharis palustris	Cyperaceae	Native	Sedge	Perennial	Monocot
Elymus glaucus	Poaceae	Native	Grass	Perennial	Monocot
Elymus spicatus	Poaceae	Native	Grass	Perennial	Monocot
Elymus trachycaulus	Poaceae	Native	Grass	Perennial	Monocot
Erodium cicutarium	Geraniaceae	Exotic	Forb	Annual, biennial	Dicot
Festuca ovina	Poaceae	Exotic	Grass	Perennial	Monocot
Galium aparine	Rubiaceae	Native	Vine, forb	Annual	Dicot
Glycyrrhiza lepidota	Fabaceae	Native	Forb	Perennial	Dicot
Gutierrezia sarothrae	Asteraceae	Native	Forb, shrub	Perennial	Dicot
Heterotheca villosa	Asteraceae	Native	Forb	Perennial	Dicot
Isatis tinctoria	Brassicaceae	Exotic	Forb	Annual, perennial	Dicot
Juncus arcticus	Juncaceae	Native	Rush	Perennial	Monocot
Juniperus communis	Cupressaceae	Native	Shrub, tree	Perennial	Gymnosperm
Lactuca serriola	Asteraceae	Exotic	Forb	Annual, biennial	Dicot
Leymus cinereus	Poaceae	Native	Grass	Perennial	Monocot
Linaria dalmatica	Scrophulariaceae	Exotic	Forb	Perennial	Dicot
Maianthemum stellatum	Liliaceae	Native	Forb	Perennial	Monocot
Medicago sativa	Fabaceae	Exotic	Forb	Annual, perennial	Dicot
Melilotus officinalis	Fabaceae	Exotic	Forb	Annual, biennial, perennial	Dicot
Mentha arvensis	Lamiaceae	Native	Forb	Perennial	Dicot
Opuntia polyacantha	Cactaceae	Native	Shrub	Perennial	Dicot
Phalaris arundinacea	Poaceae	Native	Grass	Perennial	Dicot
Plantago major	Plantaginaceae	Exotic	Forb	Perennial	Dicot
Poa bulbosa	Poaceae	Exotic	Grass	Perennial	Monocot
Poa pratensis	Poaceae	Exotic <sup>a</sup>	Grass	Perennial	Monocot
Populus spp.	Salicaceae	Native	Tree	Perennial	Dicot
Pseudoroegneria spicata	Poaceae	Native	Grass	Perennial	Monocot
Rhus trilobata	Anacardiaceae	Native	Shrub	Perennial	Dicot
Ribes aureum	Grossulariaceae	Native	Shrub	Perennial	Dicot
Rosa woodsii	Rosaceae	Native	Subshrub	Perennial	Dicot
Rumex crispus	Polygonaceae	Exotic	Forb	Perennial	Dicot
Salix spp.	Salicaceae	Native	Shrub, tree	Perennial	Dicot
Schoenoplectus acutus	Cyperaceae	Native	Grass	Perennial	Monocot
Senecio riddellii	Asteraceae	Native	Forb	Perennial	Dicot
Smilax spp.	Smilacaceae	_	Shrub	Perennial	Monocot
Solanum dulcamara	Solanaceae	Exotic	Forb, vine	Perennial	Dicot
Solidago velutina	Asteraceae	Native	Forb	Perennial	Dicot
Taraxacum officinale	Asteraceae	Exotic	Forb	Perennial	Dicot
Toxicodendron radicans	Anacardiaceae	Native	Forb, shrub, vine	Perennial	Dicot
Tragopogon dubius	Asteraceae	Exotic	Forb	Annual, biennial	Dicot
Urtica dioica	Urticaceae	Native	Forb	Perennial	Dicot

<sup>a</sup>The USDA lists the native status of *Poa pratensis* (Kentucky bluegrass) as debated; following the *Manual of Grasses for North America* (ed. Barkworth et al., 2007), we classify the species as exotic but naturalized. Barkworth, M. E., L.K. Anderton, K.M. Capels, S. Long, and M. B. Piep [eds.], 2007. Manual of Grasses for North America. Intermountain Herbarium and Utah State University Press, Logan, Utah, USA.